

EVALUATION ON THE ECOLOGICAL STATUS OF THE MACROZOOBENTHIC COMMUNITIES IN THE MARANO AND GRADO LAGOON (NORTHERN ADRIATIC SEA)

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ABSTRACT

The status of the macrozoobenthic community in the Marano and Grado Lagoon was evaluated, according to the application of the Water Framework Directive for the transitional waters. Benthos samplings were carried out in 2008. At forty-two sampled stations 14,522 organisms from 163 taxa were identified. The number of taxa and diversity indexes decreased from inlets towards the inner bank of the lagoon, as a function of a salinity gradient identified with three water types. Multivariate analysis and analysis of benthic biocenoses revealed the existence of three distinct macrozoobenthic communities related to the closeness and/or the proximity of sea. Dominant species resulted to be typical resident of lagoonal environment, accompanied with opportunistic species able to tolerate large variations of chemical and physical parameters in the transitional environments. M-AMBI index assigned the ecological quality status in relation to biodiversity degree.

Key words: macrozoobenthos, Lagoon of Marano and Grado, ecological status, Water Framework Directive, northern Adriatic Sea

VALUTAZIONE DELLO STATO ECOLOGICO DELLE COMUNITA' MACROZOOBENTONICHE NELLA LAGUNA DI MARANO E GRADO (ADRIATICO SETTENTRIONALE)

SINTESI

Lo stato ecologico delle comunità macrozoobentoniche della Laguna di Marano e Grado è stato esaminato in applicazione della Direttiva Comunitaria sulle acque superficiali per gli ambienti di transizione. Il macrozoobenthos è stato campionato nel 2008 su 42 stazioni e sono stati identificati 14.522 individui per un totale di 163 taxa. Il numero di taxa e gli indici di diversità hanno presentato valori decrescenti dalle bocche lagunari verso le aree interne della laguna, in funzione del gradiente di salinità identificato nei tre tipi idrici. L'analisi multivariata e l'analisi bioecologica hanno rilevato la presenza di tre distinte comunità bentoniche in relazione alla vicinanza e/o lontananza dal mare. Le specie dominanti sono risultate quelle tipiche degli ambienti lagunari, accompagnate da specie opportuniste capaci di tollerare l'elevata variabilità ambientale degli ambienti di transizione. L'indice M-AMBI ha evidenziato come lo stato di qualità ecologica venga attribuito in funzione del grado di biodiversità.

Parole chiave: macrozoobenthos, Laguna di Marano e Grado, stato ecologico, Direttiva Quadro sulle Acque, Adriatico Settentrionale

INTRODUCTION

Lagoons are classified as transitional waters sited between marine and continental domains. These systems are brackish or hyperhaline water bodies, separated from sea by barrier islands, formed in transgressional regime due to the presence of abundant terrigenous supplies and significant coastal transports (Brambati, 1988). Until 1950's lagoons, and generally all transitional waters, were classified according to salinity, but the tools employed were quite different. Finally in 1958 the "Venice system" was proposed as a water classification system according to fixed salinity values.

The Water Framework Directive (WFD 2000/60/CE) defines transitional waters as "*all the surface water bodies in the vicinity of river mouths which are partly saline in character as a result of their proximity to coastal waters but which are substantially influenced by fresh water flows*". According to this definition, transitional waters are all ecotones situated between terrestrial, freshwater and marine ecosystems, characterized by high spatial heterogeneity and temporal variability (Basset *et al.*, 2006a). On this basis, transitional waters include fjords, river mouths, deltas, rias, lagoons, coastal ponds and estuaries (McLusky & Elliott, 2007). Transitional waters are heterogeneous and dynamic ecosystems (Gomez *et al.*, 1998; Benedetti-Cecchi *et al.*, 2001) which morphology and hydrology change quickly under the influence of high sedimentation rates, natural coastal dynamics and frequent human activities (Ver *et al.*, 1999; Pastres *et al.*, 2004). These habitats often show high trophic fluxes, broad ranges of chemical and physical parameters with fast biogeochemical cycles (Herbert, 1999; Petihakis *et al.*, 1999; De Wit *et al.*, 2001). In addition, due to their shallow depth and scarce renewal of waters, most of the transitional ecosystems are very vulnerable to eutrophication and chemical pollution (Barnes, 1999), leading to rapid and often unpredictable changes in communities' composition and functioning (Herbert, 1999; Sfriso *et al.*, 2001; Mistri *et al.*, 2002a). The conservation and management of transitional waters requires monitoring activity, integrating chemical and physical evaluation with biological assessment (Gibson *et al.*, 2000; Logan & Furse, 2002).

In shallow water systems, such as lagoons, benthic compartment plays a crucial role controlling the main ecological processes; therefore changes in its structure could affect the whole ecosystem (Snelgrove *et al.*, 1997; Weslawski *et al.*, 2004; Tenore *et al.*, 2006). Due to this, it is possible to estimate the effects of different ecological drivers on the ecosystems' functioning, by analyzing modifications of lagoon benthic communities over time (Pranovi *et al.*, 2008).

The Marano and Grado Lagoon is a part of the lagoon system of the northern Adriatic Sea, stretching between the mouths of the Po and Isonzo rivers (Bram-

bati, 1988). The lagoon, which is located among the Isonzo river to the East and the Tagliamento river to the West, has a total surface area of 160 km², and extends parallel to the coastline along 32 km (Falace *et al.*, 2009).

Aristocle Vatova has provided most of the main studies and notions dealing with structure of benthic communities in the northern Adriatic lagoons. Since 1930 he studied hydrology, benthic flora and fauna in the Venice Lagoon (Vatova, 1940, 1949), subsequently benthic fauna and productivity of the Marano and Grado Lagoon (Vatova, 1964a, 1964b, 1965). Vatova (1964a) described the general features of the Marano and Grado lagoons, focusing on differences between the two basins. In particular the author measured a lower mean salinity in the Marano Lagoon (21‰) than the Grado Lagoon (26‰) and this difference allowed distinguishing the basins and the related distribution of benthic communities. Benthic fauna of the Marano was in fact poorer than that of Grado, both in term of species and abundance, due to higher freshwater inputs. On the other hand, it was more productive in term of biomass.

The most recent and exhaustive study on macrozoobenthos in the Marano and Grado Lagoon was carried out during a three-year study, from 1993 to 1995 (Orel *et al.*, 2001; Zamboni, 2008). Authors substantially confirmed the observations pointed out by Vatova (1964a): a decreasing gradient of biodiversity was observed moving from the Grado to the Marano, as well as from inlets to inner areas of the lagoon. Orel *et al.* (2001) and Zamboni (2008) identified the zonation of benthos on the basis of confinement degree in the paralic environments as proposed by Guelorgè & Perthuisot (1983), who defined paralic environments the aquatic ecosystems which have, or had, relation with the sea.

In 1986 the Manila clam (*Tapes philippinarum*) was introduced in the Marano Lagoon for aquaculture purposes. In the time its irrational harvesting with mechanical dredges out of farming areas impacted the benthic community (Orel *et al.*, 2002, 2005). Since 2006 the uncontrolled use of dredges was stopped and manual harvesting is permitted beyond the farming areas.

The aim of the paper was to evaluate the ecological status of the macrozoobenthic community in the Marano and Grado Lagoon, as required by the application in the WFD.

MATERIAL AND METHODS

The Marano and Grado Lagoon is defined as a coastal microtidal lagoon with large dimensions (Italian Ministry of Environment Decree n.131/08) and the related water types were established as a function of salinity values. In detail, we can discriminate between the mesohaline lagoon (5–20 psu), polyhaline lagoon (20–30 psu) and euhaline lagoon (30–40 psu). Forty-two



Fig. 1: Study area and sampling stations.
Sl. 1: Obravnavano območje in vzorčevalne postaje.

sampling sites were selected both on the basis of water types, surface and potential gradient of confinement from sea inlets to inner areas (Fig. 1).

Benthic samples were collected with a 0.047 m² van Veen grab in May 2008. At each station four grabs were taken. The sediment was sieved on a 1 mm mesh and fixed in 4% buffered formaldehyde solution stained with Bengal Rose, and then the fauna was separated and identified to the lowest possible taxonomical level.

Uni- and multivariate techniques were employed to analyze the communities' structure including: abundance, number of taxa, diversity indexes (Shannon-Wiener diversity index (H') on \log_2 basis (Shannon & Weaver, 1949), Margalef's index (d) (Margalef, 1958) and Pielou's evenness index (J) (Pielou, 1966). The Bray-Curtis similarity coefficient was calculated on square-root transformed data, using complete linkage; subsequently, one-way ANOSIM, K-dominance curves and SIMPER analysis were applied to evaluate similarity and/or differences among groups (PRIMER software package developed at the Plymouth Marine Laboratory).

Bionomic percentage affinity ($A\%$) was calculated by considering characteristic species according to Pères & Picard (1964). The correction coefficient C was first calculated as a percentage of characteristic species of bio-

cenosis j respect to the ones of other biocenosis. Then, the absolute affinity of each station was calculated as:

$$A_j = n_j (100 - C_j)$$

where n_j is the number of characteristic species of biocenosis j in the considered station. Finally, using a simple proportion, this parameter was expressed as percentage affinity ($A\%$).

$A\%$ was calculated for each biocenosis found in the lagoon: Euryhaline and Eurythermal Lagoon biocenosis (LEE French acronym for biocoenose Lagunaire Euryhaline et Eurytherme), fine well-sorted sand biocenosis (SFBC biocoenose des Sables Fins Bien Calibres), superficial muddy sand in sheltered areas biocenosis (SVMC biocoenose des Sables Vaseux Superficiels en Mode Calme), coastal terrigenous muds biocenosis (VTC biocoenose des Vases Terrigenes Cotieres), *Posidonia oceanica* meadow biocenosis (HP biocoenose de l'Herbier de Posidonies), fine superficial sand biocenosis (SFS biocoenose des Sables Fins Superficiels) and coastal detrital bottoms biocenosis (DC biocoenose des fonds Detritiques Cotiers).

The Bray-Curtis similarity coefficient was calculated on not transformed $A\%$ data, using complete linkage. Furthermore RELATE procedure was used to compare clusters derived from abundance and $A\%$ data.

Tab. 1: Average values of number of taxa, abundance, H', d, J and summary of Kruskal-Wallis one-way analysis of variance applied to the water types.

Tab. 1: Povprečne vrednosti števila taksonov, številčnosti, H', d, J in povzetek Kruskal-Wallisove enosmerne analize variance na tipih voda.

	No. taxa	Abundance (ind./m ²)	H'	d	J
Euhaline	38±14	1,775±1,773	3.79±0.68	6.46±2.00	0.74±0.10
Polyhaline	26±8	1,937±1,698	3.01±0.81	4.37±1.19	0.65±0.15
Mesohaline	13±4	1,711±892	2.13±0.50	2.06±0.70	0.59±0.09
Kruskal-Wallis (H)	18.49	0.026	16	20.2	5.23
p-value	≤0.0001	≤0.99	≤0.0003	≤0.0001	≤0.073
df	2	2	2	2	2

AMBI and Biotic Index (BI) were applied (Borja *et al.*, 2000) using the AMBI program (AZTI Marine Biotic Index) (www.azti.es). These indexes are based on the classification of the benthic species in five (I-V) ecological groups (EG), according to their tolerance to pollution (from EG-I = species very sensitive to organic enrichment, intolerant to pollution, EG-II = species indifferent to enrichment, EG-III = species tolerant to enrichment, slightly unbalanced environments, EG-IV = second-order opportunistic species, slight to pronounced unbalanced environments, to EG-V = first-order opportunistic species, pronounced unbalanced environment), then applying an algorithm to calculate the AMBI on a scale of increasing pollution (from 1 to 6) and obtaining the corresponding BI (from 0-1 = unpolluted sites, 2 = slightly polluted, 3 = moderately polluted, 4-5 = moderately to heavily polluted, 6 = heavily polluted to 7 = extremely polluted, azoic state). M-AMBI (Multivariate AMBI) was calculated to assess the ecological quality status (EcoQS): High, Good, Moderate, Poor and Bad according to WFD. This index includes the species richness, Shannon-Wiener diversity and AMBI at the very same time (Muxika *et al.*, 2007). EcoQS was evaluated on the basis of the reference conditions proposed for Italian transitional waters (ISPRA, 2010).

RESULTS

14,522 organisms from 163 taxa (142 species determined) were identified. Polychaetes were by far the dominant group with 72 species followed by the molluscs (36 species), crustaceans (25 species), echinoderms (6 species) and "other". This latter usually represents scarce groups such as ascidians, anthozoans, sipunculids, nemertines, phoronids, turbellarians and larvae of insects. Taking into consideration the three water types, 147 taxa were recorded in the euhaline water type (16 sampling stations) for a total of 5,339 individuals; 106 taxa were collected in the polyhaline water type (20 sampling stations), for a total of 7,283 individuals and finally 36 taxa in the mesohaline water type (6 sampling stations), for a total of 1,930 individuals.

The percentage abundance of groups was quite proportional in each water type, but the disappearance of echinoderms in the mesohaline lagoon was notable (Fig. 2). Table 1 shows mean values in each water type of the number of taxa, abundance, H', d and J. A clear decreasing gradient from euhaline lagoon to mesohaline basin was significant for taxa, H' and d (Tab. 1); mean abundance did not show any gradient from inlet to inner areas, whereas mean J was decreasing but not in a significant manner.

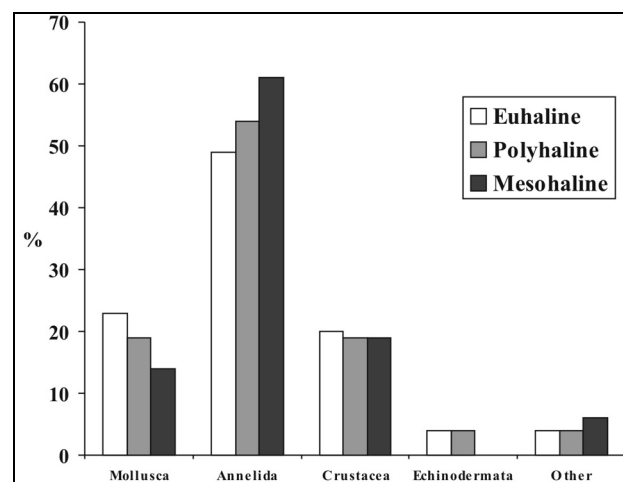


Fig. 2: Percentage of taxa for different animal phyla detected in euhaline, polyhaline and mesohaline water types.

Sl. 2: Delež taksonov za različna živalska debla v evhalinem, polihalinem in mezohalinem tipu vode.

The dendrogram resulting from the Bray-Curtis similarity matrix showed three different groups tested with one-way ANOSIM ($R=0.578$; $p<0.001$): stations close to inlets (group 1), stations among inlets and the inner bank (group 2) and stations closed to the inner bank (group 3) (Fig. 3).

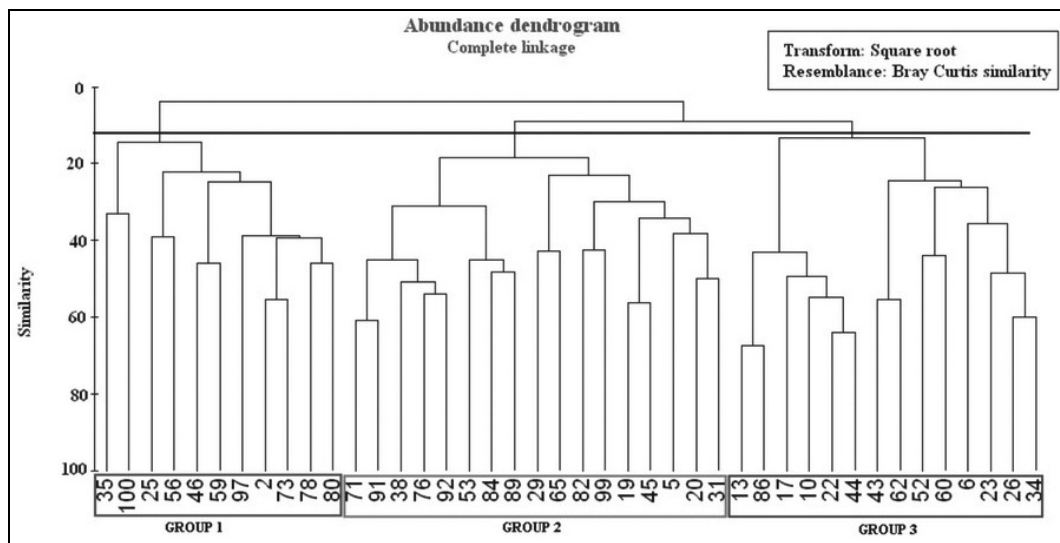


Fig. 3: Dendrogram obtained by taxa abundance values.
Sl. 3: Dendrogram številčnosti taksonov.

Tab. 2: Cumulative percentage of dominant taxa in each group revealed by SIMPER analysis.

Tab. 2: Kumulativni odstotek dominantnih taksonov v vsaki skupini, ugotovljen s SIMPER analizo.

Group	Taxa	Cum. %
Paralic group	<i>Abra segmentum</i>	34.7
	<i>Hediste diversicolor</i>	55.05
	<i>Streblospio shrubsolii</i>	66.94
Mixed group	<i>Chaetozone</i> sp.	17.75
	<i>Abra segmentum</i>	34.76
	<i>Heteromastus filiformis</i>	44.91
	<i>Streblospio shrubsolii</i>	53.34
	Oligochaeta indet.	61.48
Marine group	<i>Minuspio cirrifera</i>	12.02
	<i>Mediomastus capensis</i>	21.97
	<i>Heteromastus filiformis</i>	30.16
	<i>Abra segmentum</i>	37.55
	<i>Pseudoleiocardia fauveli</i>	44.59
	<i>Capitella capitata</i>	51.12
	Oligochaeta indet.	56.84
	<i>Myriochele oculata</i>	62.1

Trend of k-dominance curves identified the community structure in each of the identified groups: group 1, or marine group, with the highest number of species and a quite regular and homogeneous trend; group 3, or paralic group, with higher slope and the lowest number of species; group 2, or mixed group, with an intermediate shape (Fig. 4).

SIMPER analysis identified the mostly involved species in the three groups (Tab. 2). The paralic group was mainly represented by the bivalve *Abra segmentum*, the nereid polychaet *Hediste diversicolor* and the spionid *Streblospio shrubsolii*. The mixed group was character-

ized by the abundance of the capitellid *Heteromastus filiformis*, the cirratulid *Chaetozone* sp. and oligochaets, in addition to *A. segmentum* and *S. shrubsolii*. The Marine group, in addition to *A. segmentum*, *H. filiformis* and oligochaets, recorded the abundance of the capitellids *Mediomastus capensis*, *Pseudoleiocardia fauveli*, *Capitella capitata*, the spionid *Minuspio cirrifera* and the owenid *Myriochele oculata*.

Overall, 26 exclusive and preferential species for 7 biocenoses were found (Tab. 3). The dendrogram obtained from the Bray-Curtis similarity matrix applied on A% data, highlighted three groups of stations on the ba-

sis of the different A% values for Eurythermal and Euryhaline biocoenosis (LEE) (Fig. 5): the Paralic group having $A\%_{LEE} > 65\%$, the Marine group with $A\%_{LEE} < 30\%$ and Mixed group $30\% < A\%_{LEE} < 65\%$. The RELATE procedure revealed a similarity of the groups identified by two dendrograms ($Rho=0.546$; $p < 0.0001$).

Tab. 3: Characteristic species and biocoenoses detected in the study area: Euryhaline and Eurythermal Lagoon biocoenosis (LEE), fine well-sorted sand biocoenosis (SFBC), superficial muddy sand in sheltered areas biocoenosis (SVMC), coastal terrigenous muds biocoenosis (VTC), Posidonia oceanica meadow biocoenosis (HP), coastal detrital bottoms biocoenosis (DC) and fine superficial sand biocoenosis (SFS).

Tab. 3: Značilne vrste in biocenozoze v obravnavanem območju: biocenoza evrihaline in evritermne lagune (LEE), biocenoza na finem sortiranem pesku (SFBC), biocenoza zamuljenih peskov v zaščitenih predelih (SVMC), biocenoza obalnega terigenega mulja (VTC), biocenoza podvodnih travnikov Posidonia oceanica (HP), biocenoza obalnega detritičnega dna (DC) in biocenoza finega površinskega peska (SFS).

Characteristic species	Biocoenosis
<i>Gibbula adriatica</i>	LEE
<i>Cerastoderma glaucum</i>	
<i>Abra segmentum</i>	
<i>Tapes philippinarum</i>	
<i>Hediste diversicolor</i>	
<i>Streblospio shrubsolii</i>	
<i>Carcinus mediterraneus</i>	
<i>Nephtys hombergi</i>	SFBC
<i>Owenia fusiformis</i>	
<i>Prionospio caspersi</i>	
<i>Diogenes pugilator</i>	
<i>Tellimya ferruginosa</i>	
<i>Thracia papyracea</i>	
<i>Euclymene oerstedii</i>	
<i>Upogebia pusilla</i>	SVMC
<i>Loripes lacteus</i>	
<i>Tapes decussates</i>	
<i>Paphia aurea</i>	
<i>Petaloproctus terricolus</i>	VTC
<i>Sternaspis scutata</i>	
<i>Ampharete acutifrons</i>	
<i>Laonice cirrata</i>	HP
<i>Venus verrucosa</i>	
<i>Euclymene lumbricoides</i>	DC
<i>Abra prismatica</i>	
<i>Glycera tridactyla</i>	SFS

The EcoQS assigned with M-AMBI, was Good for 44% of euhaline stations, Moderate for 38%, Poor for 12% (st. 76 and 100) and Bad for 6% (st. 99) (Fig. 6). In polyhaline type the stations with High status corresponded to 20%, Good to 50%, Moderate to 25% and Poor to 5% (st. 86) (Fig. 6). In the mesohaline type one station (st. 43) was Good, three stations were assessed as Moderate (50%) and two were Poor (st. 13 and 17) (Fig. 6).

The summarizing Table 4 reports M-AMBI as EcoQS, AMBI as disturbance classification and the groups (Marine, Mixed, Paralic) assigned with clusters. The mean number of taxa and average Shannon-Wiener index (H') were calculated for each EcoQS of the water types.

Within the euhaline type (salinity 30-40 psu) all stations with Good ecological quality status possessed Marine characteristics, with 50 identified taxa and $H' > 4$. Stations with Moderate EcoQS possessed normally Mixed characteristics, on average 29 taxa and mean $H'=3.5$. Stations having Poor EcoQS possessed 33 taxa and $H'=2.9$: st. 100 had Marine characteristics but a Poor status was assigned probably because of high dominance of oligochaetes, indicating strongly unbalanced conditions. Bad EcoQS was assigned to st. 99, which was very poor in term of taxa. In the polyhaline type (salinity 20-30 psu) only two stations have Marine characteristics (st. 25 and 35), but normally stations showed Mixed or sometimes Paralic conditions. Stations with High EcoQS possessed on average 38 taxa and $H'=4$. Stations with Good EcoQS had 24 taxa and $H' > 3$. In mesohaline type (salinity 5-20 psu), stations have Paralic conditions, except for st. 43 which is Mixed-Paralic with Good EcoQS (17 taxa, $H'=2.5$). Disturbance classification detected with AMBI varied from slightly disturbed to moderately disturbed and it seems not linked to water type and cluster groups.

DISCUSSION

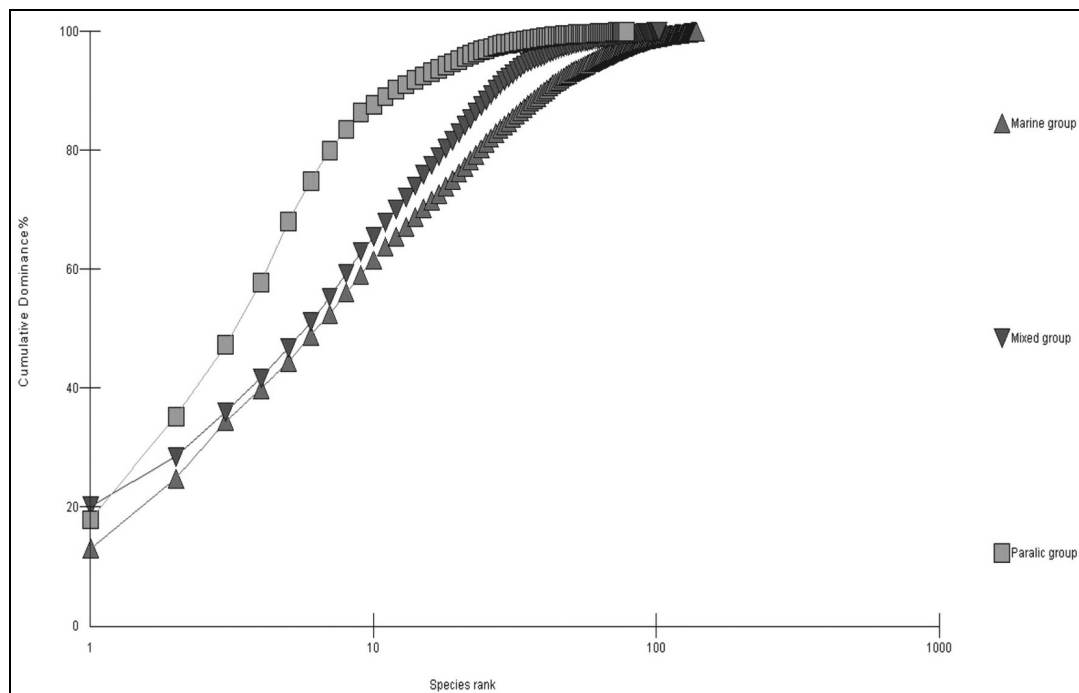
The aim of this study was to characterize the macrozoobenthos living in the Marano and Grado Lagoon, 15 years after the studies carried out in 1993–1995 (Drioli, 1995/1996; Zamboni, 1995/1996; Orel *et al.*, 2001; Zamboni, 2008), and to provide a preliminary application of the WFD 2000/60/CE in this lagoon.

Previously, several authors classified lagoons and coastal ponds as a function of salinity values (Redeke, 1922, 1933; Brunelli, 1933; D'Ancona, 1959). The salinity is a key parameter affecting the biological organization and the expression of biodiversity in Italian lagoons (Basset *et al.*, 2006b) and Greek brackish environments (Reizopoulou & Nicolaidou, 2004). Three water types were identified and analyzed in Marano and Grado according to the actual Italian legislations.

Tab. 4: Summarizing table.

Tab. 4: Pregledna tabela.

sampling station	water type	EcoQS	No. taxa	H'	groups cluster	groups cluster	AMBI
		(M-AMBI)	(mean values)	(mean values)	A%	abundance	disturbance classification
46	EU	GOOD	50	4.4	Marine	Marine	Slightly
56	EU				Marine	Marine	Slightly
59	EU				Marine	Marine	Slightly
73	EU				Marine	Marine	Moderately
78	EU				Mixed	Marine	Moderately
80	EU				Marine	Marine	Moderately
84	EU				Marine	Mixed	Slightly
29	EU	MODERATE	29	3.5	Mixed	Mixed	Slightly
38	EU				Mixed	Mixed	Moderately
52	EU				Marine	Paralic	Slightly
89	EU				Mixed	Mixed	Moderately
92	EU				Mixed	Mixed	Moderately
97	EU				Mixed	Marine	Slightly
76	EU	POOR	33	2.9	Mixed	Mixed	Moderately
100	EU				Marine	Marine	Moderately
99	EU	BAD	14	3	Mixed	Mixed	Moderately
2	POLY	HIGH	38	4	Paralic	Marine	Moderately
23	POLY				Paralic	Paralic	Slightly
31	POLY				Mixed	Mixed	Slightly
53	POLY				Paralic	Mixed	Slightly
5	POLY	GOOD	24	3.2	Mixed	Mixed	Slightly
19	POLY				Mixed	Mixed	Moderately
20	POLY				Mixed	Mixed	Moderately
26	POLY				Paralic	Paralic	Moderately
34	POLY				Mixed	Paralic	Slightly
45	POLY				Mixed	Mixed	Moderately
60	POLY				Paralic	Paralic	Slightly
62	POLY				Mixed	Paralic	Slightly
65	POLY				Mixed	Mixed	Moderately
71	POLY				Mixed	Mixed	Moderately
6	POLY	MODERATE	22	2.2	Paralic	Paralic	Moderately
25	POLY				Marine	Marine	Slightly
35	POLY				Marine	Marine	Moderately
82	POLY				Paralic	Mixed	Moderately
91	POLY				Mixed	Mixed	Moderately
86	POLY	POOR	10	1.6	Paralic	Paralic	Slightly
43	MESO	GOOD	17	2.5	Mixed	Paralic	Slightly
10	MESO	MODERATE	14	2.4	Paralic	Paralic	Slightly
22	MESO				Paralic	Paralic	Slightly
44	MESO				Paralic	Paralic	Moderately
13	MESO	POOR	9	1.6	Paralic	Paralic	Slightly
17	MESO				Paralic	Paralic	Slightly



**Fig. 4: K-dominance curves relative to three different groups detected by dendrogram.
Sl. 4: Krivulje K-dominance za tri različne skupine, ugotovljene z dendrogramom.**

The Marano Lagoon is embedded in the inland, coupled to a less marine water exchange and very abundant freshwater inputs from Stella, Corno and Aussa rivers (Fig. 1); on the contrary the Grado Lagoon is characterized by scarce freshwater inputs and a wider exchange of marine water. The subdivision of the whole basin into three water types reflects quite well the previous observations of Vatova (1964a, 1964b, 1965) and more recently those of Orel *et al.* (2001) and Zamboni (2008). During 2008 taxa richness, diversity and evenness indices showed high values particularly in nearby inlets. In these areas Shannon-Wiener index (H') was >4 , which can be considered a very high value for soft bottom macrozoobenthic communities (Gray, 2000). Therefore high indices values and relevant number of species recorded in 2008 (163 taxa) compared to the 90's, (85 taxa in the three years; Zamboni, 2008), could indicate the increasing richness of taxa in the Marano and Grado Lagoon during the last 15 years. The number of taxa decreased moving from the inlets towards the inner bank of the lagoon, as described also in previous studies. On the contrary, abundance did not show any distribution gradient. On the whole, the partition of benthos into three salinity bands was confirmed by a decreasing diversity from the inlets towards the inner bank. Due to this, a monitoring program based on water types is fully supported by observations carried out in the first studies (Vatova, 1964a, 1965; Orel *et al.*, 2001; Zamboni, 2008) and the 2008 survey.

Guelorgèt & Perthuisot (1982, 1983, 1992) referred on the existence of peculiar euryhaline species. This feature was previously emphasized by Pérès & Picard (1964), who defined the Euryhaline and Eurythermal Lagoon biocenosis. This biocenosis is typical of unstable environments where higher or lower salinity and/or wider salinity and temperature variations occur during the year. This peculiarity is mainly due to river floods, rainfalls and high summer rate of evaporation. Moreover, living communities are able to recover rather quickly their original structures when environmental disturbances and/or dystrophic crisis occur. Due to this great resilience, rather than stability of the communities, paralic environments prevail. The previously mentioned peculiar characteristics of the communities led Guelorgèt & Perthuisot (1982, 1983, 1992) to divide the paralic domain into two subsets. The first one as near paralic, close to the inlets and with chemical-physical parameters similar to sea, whereas the second one as far paralic, farther from sea with chemical-physical parameters deeply different to marine domain. As previously discussed, the salinity is considered a key parameter determining a transition gradient from a typical marine community to a freshwater one. In fact, the fresh/marine waters exchange and circulation play an important role in the dynamics of a lagoon. In this way, hydrology is the fundamental factor conditioning any lagoon environment. In the same context Guelorgèt & Perthuisot (1992) defined the confinement as the time of renewal of

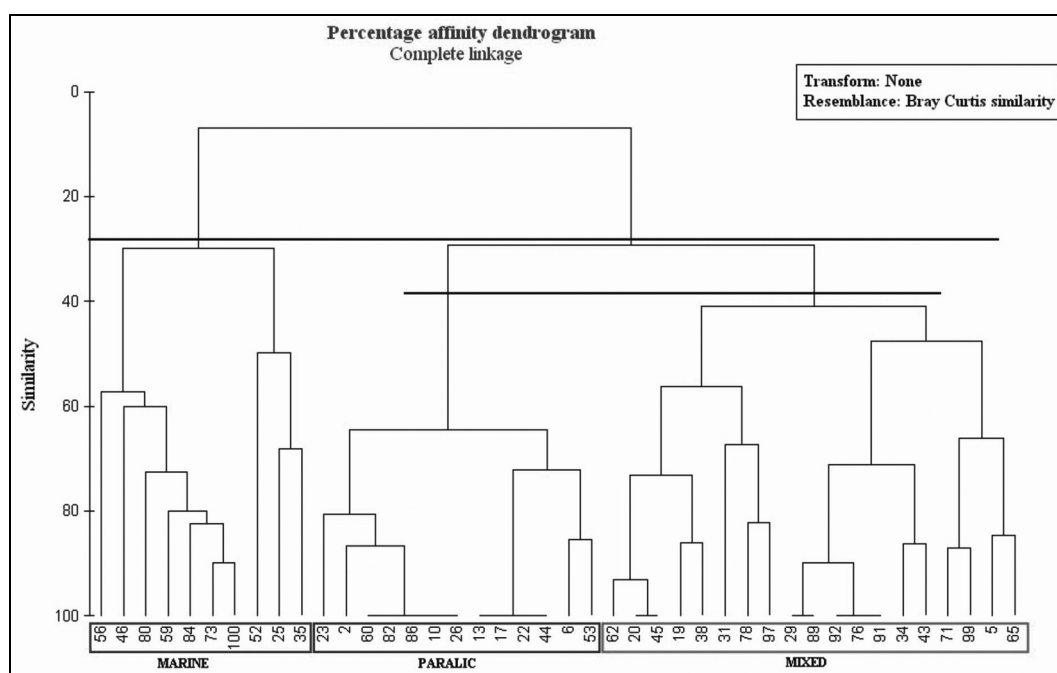


Fig. 5: Dendrogram obtained by percentage affinity values.
Sl. 5: Dendrogram, narejen na podlagi afinitetnih vrednosti (v %).

marine elements in relation to the extension of paralic domain. However, this definition should be considered on a larger scale since it is dependent on many other parameters such as tidal range, width inlets, freshwaters inputs, wind regime, depth, trophic conditions, transparency, oxygen content and so on.

Multivariate analysis in 2008 defined three different areas as a function of closeness to inlets and freshwaters inputs. The group of stations close to inlets could be defined as a lagoon community having marine characteristics; stations nearby the inner bank as a strictly paralic community; finally stations showing a combination of marine and paralic properties as a mixed community. This kind of partition was confirmed and validated by ANOSIM analysis; moreover k-dominance curves focused a basic community structure in each group. In the same way SIMPER analysis punctually identified the mostly involved species of the mentioned groups.

Three characteristic species of paralic environments and LEE biocenosis represent at least 60% of relative abundance in the Paralic group and dominate in inner zones where marine species cannot survive: *Abra segmentum*, *Hediste diversicolor* and *Streblospio shrubsolii*. *A. segmentum* is typically euryhaline, frequent in oligo-hyperhaline waters, tolerates a wide range of salinities, from 3 to 41 psu (Marazanof, 1969; Kevrekidis, 2004). It is common and frequently abundant in the Mediterranean coastal lagoons, where it plays a dominant role, both in terms of number and biomass, in the infauna of these habitats and represents an important food for fish

(Kevrekidis & Kasapis, 2009). *H. diversicolor* preferentially lives in muddy sediments. It is a typical inhabitant of European brackish water habitats and shows a high level of tolerance towards different types and concentrations of contaminants (Volpi *et al.*, 1999). *H. diversicolor* tolerates a wide salinity range, preferring, in any case, areas characterized by low salinity values (Guerzoni & Tagliapietra, 2006) and it is largely distributed in Mediterranean lagoons (Bazaïri *et al.*, 2003; Nicolaidou *et al.*, 2005). *S. shrubsolii* is a typical lagoon species widely distributed also in other Mediterranean lagoons (Mistri *et al.*, 2002b; Rossi & Lardicci, 2002; Bazaïri *et al.*, 2003; Dauer *et al.*, 2003). Furthermore, Mixed and Marine groups were characterized by the dominance of polychaetes belonging to Capitellidae and Spionidae families represented by small size species capable to settle in habitats having a strong variability, such as transitional environments (Holte & Oug, 1996; Mistri *et al.*, 2002b) and adapted to colonize habitats with organic enrichment (Mistri *et al.*, 2001; Thouzeau *et al.*, 2007). Oligochaets, indicating desalinized water (Nicolaidou *et al.*, 2005), were abundant, too. In addition, the owenid *Myriochele oculata*, a sandy species especially diffused nearby inlets, contributed to define the Marine group.

The bionomic analysis (Pères & Picard, 1964) revealed the existence of seven biocenoses. The main were LEE and SFBC, this latter is typical of marine domain and largely diffused in Mediterranean Sea as a soft bottom infralittoral biocenosis. The second dendrogram obtained by A% values was comparable to abundance

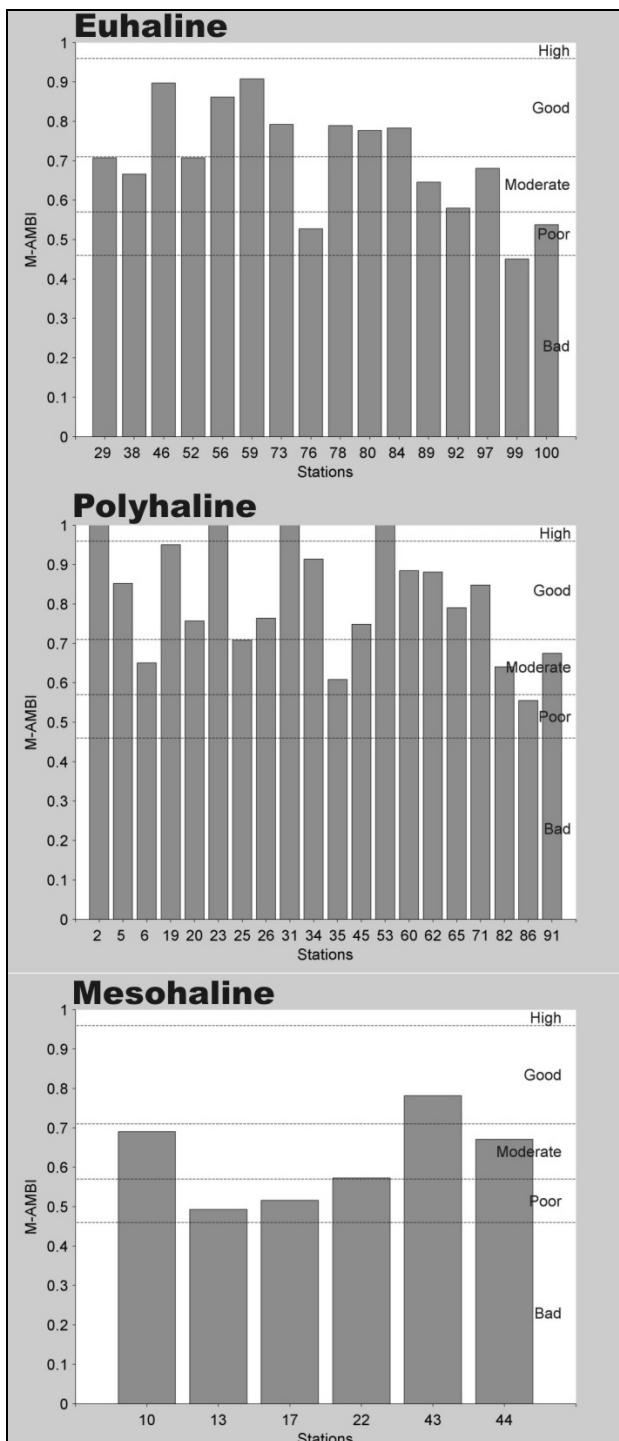


Fig. 6: EcoQS of euhaline, polyhaline and mesohaline stations according to M-AMBI index.

Sl. 6: Ekološko stanje kakovosti (EcoQS) za evhaline, polihaline in mezohaline postaje na podlagi izračuna indeksa M-AMBI.

cluster as revealed by RELATE procedure. In this way stations having Marine affinities are mostly represented by SFBC, whereas LEE is meanly <30%. LEE affinity in Paralic stations exceeds 65%, and LEE affinity in Mixed group ranges between 30 and 65%.

The Marine, Mixed and Paralic groups could correspond to zones II, III and IV-V respectively, as defined by Guelorgè & Perthuisot's benthic zonation (Guelorgè & Perthuisot, 1982, 1983, 1992). In fact, Marine group resembles the zone II, as a corresponding entrance into lagoon domain with the presence of most tolerant marine species belonging to SFBC and Fine Superficial Sands (SFS) biocenosis. Mixed group was similar to zone III with a remarkable scarcity or even disappearance of echinoderms. Vatova (1964a, 1965) noticed that the holothurioid *Trachythyone elongata* is the only capable to withstand sudden salinity changes and, in fact, only 7 specimens of 4 echinoderms species were found in the polyhaline water types (*T. elongata*, *Asterina gibbosa*, *Amphiura chiajei* and *Amphipholis squamata*). Finally, Paralic group resembled zone IV, because of the total disappearance of most tolerant marine species, such as echinoderms, and an absolute dominance of strictly paralic species such as *A. segmentum*, *Cerastoderma glaucum* and *H. diversicolor*. In some limited areas Paralic group resembled also zone V because of the appearance of chironomids larvae.

No specimens of the characteristic lagoon bivalve *Scrobicularia plana* were found in 2008. This species already recorded in the Marano and Grado Lagoon by Vatova (1964a, 1964b, 1965) and considered as particularly sensitive to river floods, was scarcely present in the 70's because of pollution phenomena (Battaglia *et al.*, 1972) and now it results to be absent since the 90's (Zamboni, 2008).

Macrozoobenthos in 2008 embraced a distribution comparable to the three years study 1993-1995 (Orel *et al.*, 2001; Zamboni, 2008). In the lagoon the salinity constitutes a sort of barrier to strictly marine species, leading to the selection of more tolerant organisms. In the inner part of the lagoon there is a clear dominance of LEE species accompanied by strongly opportunistic annelids.

The final goal was the application of indices proposed by WFD, in order to define the ecological quality status in each sampling station. M-AMBI showed a significant Good-Moderate quality status in euhaline stations. However, some sites located within heavily modified water bodies showed a Poor or Bad status. In polyhaline type 70% of stations showed Good to High EcoQS. Mesohaline type, the inner part of the Marano Lagoon, showed almost a Moderate status, due to natural selectivity of paralic habitats, characterized by low biodiversity. As suggested by Dauvin *et al.* (2009), the main issue in the application of indexes for the classification is that in order to determine anthropogenic stress,

they account to relative abundances of stress-tolerant species. However, these latter may also be tolerant to natural stressors. Moreover, due to the high variability of both physical and chemical parameters, transitional environments are generally characterized by low benthic diversity if compared to the marine environment with several opportunistic and tolerant organisms, withstanding these conditions. AMBI provides ecological group assignments based on expert opinion for a lot of taxa. Moreover the ecological behaviour of several species is quite different when they have to adapt to lagoonal conditions (Cognetti & Maltagliati, 2008) or to different geographical areas. The reference conditions for the final EcoQS are established by the legislation. Actually, in Italy, a draft document proposed by Italian Environmental Ministry provides the same reference condition for oligo-, meso- and polyhaline waters. As aforementioned the Marano and Grado Lagoon shows a quite different benthic community in the mesohaline

waters compared to polyhaline waters. Hence the application of a single reference condition for these different types causes to classify mesohaline waters with Moderate or Poor quality. In this situation, in order to achieve at least the Good EcoQS in the whole Marano and Grado Lagoon, macrozoobenthic community should reach a number of species comparable to a typical soft bottom community of marine environments, thus losing the characteristics and ecological functions of paralic environments.

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OCENA EKOLOŠKEGA STATUSA ZDRUŽB VODNIH NEVRETEŃČARJEV V MARANSKI IN GRADEŠKI LAGUNI (SEVERNI JADRAN)

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POVZETEK

Ocenjenjen je bil status združb vodnih nevretenčarjev v Maranski in Gradeški laguni, v skladu z aplikacijo Okvirne direktive o vodah za somornice. Vzorčenja bentosa so bila izvedena leta 2008. Na 42 vzorčevalnih postajah je bilo identificiranih 14.522 organizmov, pripadajočih 163 taksonom. Število taksonov in indeksi pestrosti so se zmanjševali od mesta dotokov proti notranjemu bregu lagune, in sicer kot funkcija gradienta slanosti treh tipov vode. Multivariatna analiza in analiza bentoških biocenoz sta razkrili obstoj treh združb vodnih nevretenčarjev, vezanih na bližino/oddaljenost od morja. Dominantne vrste so tipični prebivalci lagunskega okolja, spremljajo pa jih oportunistične vrste, ki lahko tolerirajo velike variacije kemičnih in fizikalnih parametrov v somornicah. Indeks M-AMBI opredeljuje ekološko stanje glede na biotsko raznolikost.

Ključne besede: vodni nevretenčarji, Maranska in Gradeška laguna, ekološki status, Okvirna direktiva o vodah, severni Jadran

REFERENCES

- Barnes, R. S. K. (1999):** What determines the distribution of coastal hydrobiid mudsnails within north-western Europe. *Mar. Ecol. P.S.Z.N.I.*, 20, 97–110.
- Basset, A., L. Sabetta, A. Fonnesu, D. Mouillot, T. Do Chi, P. Viaroli, G. Giordani, S. Reizopoulou, M. Abbiati & G. C. Carrada (2006a):** Typology in Mediterranean transitional waters: new challenger and perspective. *Aquat. Conserv.: Mar. Freshw. Ecosyst.*, 16, 441–455.
- Basset, A., N. Galuppo & L. Sabetta (2006b):** Environmental heterogeneity and benthic macroinvertebrate guilds in Italian lagoons. *Transit. Waters Bull.*, 1, 48–63.
- Battaglia, B., A. Brambati & G. Modena (1972):** Relazione sulla laguna di Marano-Grado: insediamenti industriali e turbamenti ecologici. Ass. reg. industria e commercio, Trieste.
- Bazairi, H., A. Bayed, M. Glémarec & C. Hily (2003):** Spatial organisation of macrozoobenthic communities in response to environmental factors in a coastal lagoon of the NW African coast (Merja Zerga, Morocco). *Oceanol. Acta*, 26, 457–471.
- Benedetti-Cecchi, L., F. Rindi, I. Bertocci, F. Bulleri & F. Cinelli (2001):** Spatial variation in development of epibenthic assemblage in a coastal lagoon. *Estuar. Coast. Shelf Sci.*, 52, 659–668.
- Borja, A., J. Franco & V. Pérez (2000):** A marine biotic index to establish the ecological quality of soft-bottom benthos within European estuarine and coastal environments. *Mar. Pollut. Bull.*, 40(12), 1100–1114.
- Brambati, A. (1988):** Lagune e stagni costieri: due ambienti a confronto. In: Carrada, G. C., F. Cicogna & E. Fresi (eds.): *Le lagune costiere: ricerca e gestione*. CLEM pubbl., Massa Lubrense (Na), pp. 9–33.
- Brunelli, G. (1933):** Ricerche sugli stagni costieri litoranei. *R.C. Accad. Lincei, Cl. Sci. Fis. Mat. Nat.*, ser. 6, vol. XVII, 1 sem., fasc. 3, 246–249.
- Cognetti, G. & F. Maltagliati (2008):** Perspectives on the ecological assessment of transitional waters. *Mar. Pollut. Bull.*, 56, 607–608.
- D’Ancona, U. (1959):** The classification of brackish waters with reference to the north Adriatic lagoons. *Arch. Oceanogr. Limnol.*, vol. XV (suppl.), 93–109.
- Dauer, D. M., H. K. Mahon & R. Sardá (2003):** Functional morphology and feeding behavior of *Streblospio benedicti* and *S. shrubsolei* (Polychaeta: Spionidae). *Hydrobiologia*, 496, 207–213.
- Dauvin, J. C., G. Bachelet, A. L. Barille, H. Blanchet, X. de Montaudouin, N. Lavesque & T. Ruellet (2009):** Benthic indicators and index approaches in the three main estuaries along the French Atlantic coast (Seine, Loire and Gironde). *Mar. Ecol.*, 30, 228–240.
- De Wit, R., L. J. Stal, B. A. Lomstein, R. A. Herbert, H. Van Gernerden, P. Viaroli, V. U. Cecherelli, F. Rodriguez-Valera, M. Bartoli, G. Giordani, R. Azzoni, B. Schaub, D. T. Welsh, A. Donnelly, A. Cifuentes, J. Anton, K. Finster, L. B. Nielsen, A. G. U. Pedersen, A. T. Neubauer, M. A. Colangelo & S. K. Hejls (2001):** Robust: the role of buffering capacities in stabilizing coastal lagoon ecosystems. *Cont. Shelf Res.*, 21, 2021–2041.
- Drioli, A. (1995/96):** Struttura ed evoluzione dei popolamenti bentonici della Laguna di Grado e Marano. Tesi in Scienze Biologiche. Università degli Studi di Trieste A.A. 1995–96, 53 p.
- Falace, A., A. Sfriso, D. Curiel, G. Mattassi & F. Aleffi (2009):** The Marano and Grado Lagoon. In: Cecere, E., A. Petrocelli, G. Izzo & A. Sfriso (eds.): *Flora and Vegetation of the Italian Transitional Water Systems*, pp. 1–16.
- Gibson, G. R., M. L. Bowman, J. Gerritsen & B. D. Snyder (2000):** Estuarine and coastal marine waters: bioassessment and biocriteria technical guidance. EPA 822-B-00-024. U.S. Environmental Protection Agency, Office of Water, Washington, DC.
- Gomez, E., B. Millet & B. Picot (1998):** Nutrient accumulation in a lagoon sediment relating to hydrodynamic conditions. *Oceanol. Acta*, 21, 805–817.
- Gray, J. (2000):** The measurement of marine species diversity, with an application to the benthic fauna of the Norwegian continental shelf. *J. Exp. Mar. Biol. Ecol.*, 250, 23–49.
- Guelorguèt, O. & J. P. Perthuisot (1982):** Structure et évolution des peuplements benthiques en milieu paralique. Comparaison entre un modèle dessalé (l’Etang du Prevost, France) et un modèle sursalé (la Bahiret el Biban, Tunisie). *Conséquences biologiques et géologiques*. *Journ. Rech. Oceanogr.* VII, 2.3.4, 2–11.
- Guerlogèt, O. & J. P. Perthuisot (1983):** Le domaine paralique. Expression géologiques, biologiques et économiques du confinement. *Travaux du laboratoire de géologie*, 16. Presses de l’Ecole Normale Supérieure, Paris, 129 p.
- Guerlogèt, O. & J. P. Perthuisot (1992):** Paralique ecosystem. Biological organization and functioning. *Vie Milieu*, 42(2), 215–251.
- Guerzoni, S. & D. Tagliapietra (2006):** Atlante della laguna. Venezia tra terra e mare. Marsilio Ed., 244 p.
- Herbert, R. A. (1999):** Nitrogen cycling in coastal marine ecosystems. *FEMS Microbiol. Rev.*, 23, 563–590.
- Holte, B. & E. Oug (1996):** Soft-bottom macrofauna and responses to organic enrichment in the subarctic waters of Tromsø, northern Norway. *J. Sea Res.*, 36(3–4), 227–237.
- ISPRA (2010):** Implementazione della direttiva 2000/60/CE: classificazione dello stato ecologico dei corpi idrici delle acque marine costiere e di transizione. MCW-TW Sistema di classificazione ecologica, 30 p.
- Kevrekidis, T. (2004):** Seasonal variations of the macrozoobenthic community structure at low salinities in a Mediterranean lagoon (Monolimni Lagoon, Northern Aegean). *Int. Rev. Hydrobiol.*, 89, 407–425.

- Kevrekidis, T. & K. Kasapis (2009):** Life cycle, population dynamics, growth and production of *Abra segmentum* (Mollusca, Bivalvia) at low salinities in a Mediterranean lagoon. *Helgol. Mar. Res.*, 63, 277–285.
- Logan, P. & M. Furse (2002):** Preparing for the European Water Framework Directive – making the links between habitat and aquatic biota. *Aquat. Conserv.: Mar. Freshw. Ecosyst.*, 12, 425–437.
- Marazanof, F. (1969):** Contribution à l'étude écologique des mollusques des eaux douces et saumâtres de Camargue. I. Milieux-especes. *Ann. Limnol.*, 5(3), 201–323.
- Margalef, D. R. (1958):** Information theory in ecology. *Gen. Syst.*, 3, 36–71.
- McLusky, D. S. & M. Elliot (2007):** Transitional waters: A new approach, semantics or just muddying the waters? *Estuar. Coast. Shelf Sci.*, 71, 359–363.
- Mistri, M., R. Rossi & E. A. Fano (2001):** Structure and secondary production of a soft bottom macrobenthic community in a brackish lagoon (Sacca di Goro, north-eastern Italy). *Estuar. Coast. Shelf Sci.*, 52, 605–616.
- Mistri, M., E. A. Fano, F. Ghion & R. Rossi (2002a):** Disturbance and community pattern of Polychaetes inhabiting Valle Magnavacca (Valli di Comacchio, Northern Adriatic Sea, Italy). *Mar. Ecol.*, 23(1), 31–49.
- Mistri, M., F. Ghion, S. Modugno & R. Rossi (2002b):** Response of macrobenthic communities to a hydraulic intervention in an enclosed lagoon (Vale di Gorino, Northern Italy). *J. Mar. Biol. Assoc. U.K.*, 81, 339–340.
- Muxika, I., A. Borja & J. Bald (2007):** Using historical data, expert judgment and multivariate analysis in assessing reference conditions and benthic ecological status, according to the European Water Framework Directive. *Mar. Pollut. Bull.*, 55, 16–29.
- Nicolaidou, A., S. Reizopoulou, D. Koutsoubas, S. Orfanidis & T. Kevrekidis (2005):** Biological components of Greek lagoonal ecosystems: an overview. *Med. Mar. Sci.*, 6(2), 31–50.
- Orel, G., R. Zamboni, F. Grimm & A. Zentilin (2001):** Evoluzione dei popolamenti bentonici della laguna di Marano e Grado (Adriatico settentrionale) in un triennio di ricerche. *Biol. Mar. Medit.*, 8(1), 424–431.
- Orel, G., G. Fontolan, I. Burla, R. Zamboni, A. Zentilin & G. Pessa (2002):** Aspetti dell'impatto della pesca della vongola verace filippina (*Tapes philippinarum*) con draghe al traino nella Laguna di Marano (Adriatico Settentrionale). *Biol. Mar. Medit.*, 9(1), 129–137.
- Orel, G., R. Zamboni & A. Zentilin (2005):** Impatto della pesca e della coltura di *Tapes philippinarum* sui fondali delle lagune alto adriatiche. In: Boatto, V. & M. Pellizzato (eds.): La filiera della vongola – *Tapes philippinarum* in Italia. Franco Angeli Ed., pp. 45–57.
- Pastres, R., C. Solidoro, S. Ciavatta, A. Petrizzo & G. Cossarini (2004):** Long-term changes of inorganic nutrients in the Lagoon of Venice (Italy). *J. Mar. Syst.*, 51, 179–189.
- Pérès, J. & J. Picard (1964):** Nouveau Manuel de Biologie Benthique de la Mer Méditerranée. *Réc. Trav. Stat. Mar. Endoume*, 31, 5–137.
- Petihakis, G., G. Triantafyllou, D. Koutsoubas, I. Allen & C. Dounas (1999):** Modelling the annual cycles of nutrients and phytoplankton in a Mediterranean lagoon (Gialova, Greece). *Mar. Environ. Res.*, 48, 37–58.
- Pielou, E. C. (1966):** The measurement of diversity in different types of biological collections. *J. Theor. Biol.*, 13, 131–144.
- Pranovi, F., F. Da Ponte & P. Torricelli (2008):** Historical changes in the structure and functioning of the benthic community in the lagoon of Venice. *Estuar. Coast. Shelf Sci.*, 76, 753–764.
- Redeke, H. C. (1922):** Zur Biologie der niederländischer Brackwassertypen (Ein Beitrag zur regionalen Limnologie). *Bijdr. Dierk.*, Amsterdam, vol. XXII, 329–335.
- Redeke, H. C. (1933):** Über den jetzigen Stand unserer Kenntnisse des Flora und Fauna des Brackwassers. *Verh. Internat. Verein. Limnol.*, VI(I), 46–61.
- Reizopoulou, S. & A. Nicolaidou (2004):** Benthic diversity of coastal brackish-water lagoons in western Greece. *Aquat. Conserv.: Mar. Freshw. Ecosyst.*, 14, 93–102.
- Rossi, F. & C. Lardicci (2002):** Role of the nutritive value of sediment in regulating population dynamics of the deposit-feeding polychaete *Streblospio shrubsolii*. *Mar. Biol.*, 140, 1129–1138.
- Sfriso, A., T. Birkemeyer & P. F. Ghetti (2001):** Benthic macrofauna changes in areas of Venice lagoon populated by seagrasses or seaweeds. *Mar. Environ. Res.*, 52, 323–349.
- Shannon, C. E. & W. Weaver (1949):** The mathematical theory of communication. Urbana, University of Illinois Press, 117 p.
- Snelgrove, P. V. R., T. H. Blackburn, P. Hutchings, D. Alongi, J. F. Grassle, H. Hummel, G. King, I. Koike, P. J. D. Lamshead, N. B. Ramsing & V. Solis-Weiss (1997):** The importance of marine biodiversity in ecosystem processes. *Ambio*, 26, 578–583.
- Tenore, K. R., R. N. Zajac, J. Terwin, F. Andrade, J. Blanton, W. Boynton, D. Carey, R. Diaz, A. F. Holland, E. Lopez-Jamar, P. Montagna, F. Nichols, R. Rosenberg, H. Queiroga, M. Sprung & R. B. Whitlatch (2006):** Characterizing the role benthos plays in large coastal seas and estuaries: a modular approach. *J. Exp. Mar. Biol. Ecol.*, 330, 392–402.
- Thouzeau, G., J. Grall, J. Clavier, L. Chauvaud, F. Jean, A. Leynaert, S. Longphurt, E. Amice & D. Amouroux (2007):** Spatial and temporal variability of benthic biogeochemical fluxes associated with macrophytic and macrofaunal distributions in the Thau lagoon (France). *Estuar. Coast. Shelf Sci.*, 72, 432–446.
- Vatova, A. (1940):** Le zoocenosi della Laguna Veneta. *Thalassia*, vol. III, 3–28.
- Vatova, A. (1949):** Caratteri di alcune facies bentoniche della Laguna Veneta. *Nova Thalassia*, vol. I(4), 3–15.

- Vatova, A. (1964a):** Ricerche quantitative sulla fauna bentonica delle lagune di Marano e Grado. Boll. Pesca Piscic. Idrobiol., 18, 3–11.
- Vatova, A. (1964b):** Sulla produttività della Laguna di Grado e Marano. Accademia Nazionale dei Lincei. Estratto dei Rendiconti della Classe di Scienze Fisiche, Matematiche e Naturali, ser. VIII, vol. XXXVII, fasc. 5, 330–333.
- Vatova, A. (1965):** Nouvelles recherches sur la faune benthique de la lagune de Grado-Marano. Rapp. p. v. Reun. Comm. Int. Explor. Sci. Mer Médit.
- Ver, L. M. B., F. T. Mackenzie & A. Lerman (1999):** Carbon cycle in the coastal zone: effects of global perturbations and change in the past three centuries. Chem. Geol., 159, 283–304.
- Volpi, G. A., L. Cavallini, E. Delaney, D. Tagliapietra, P. F. Ghetti, C. Bettiol & E. Argese (1999):** *H. diversicolor*, *N. succinea* and *P. cultrifera* (Polychaeta: Nereididae) as bioaccumulators of cadmium and zinc from sediments: preliminary results in the Venetian lagoon (Italy). Toxicol. Environ. Chem., 71, 457–474.
- Weslawski, J. M., P. V. R. Snelgrove, L. A. Levin, M. C. Austen, R. T. Kneib, T. M. Iliffe, J. R. Carey, S. J. Hawkins & R. B. Whitlatch (2004):** Marine sedimentary biota as providers of ecosystem goods and services. In: Wall, D.H. (ed.): Sustaining Biodiversity and Ecosystem Services in Soils and Sediments. Island Press, Washington, pp. 73–98.
- Zamboni, R. (1995/1996):** I popolamenti bentonici della Laguna di Grado e Marano. Tesi in scienze Biologiche. Università degli Studi di Trieste, A.A. 1995–96, 40 p.
- Zamboni, R. (2008):** Proposta di realizzazione di una cartografia bionomica come strumento di monitoraggio e di gestione della Laguna di Marano e Grado. XIX ciclo dei dottorati di ricerca in metodologie di biomonitoraggio dell'alterazione ambientale. Dipartimento di Biologia, Università degli Studi di Trieste, 185 p.